Understanding changes in bird populations – the role of bird marking

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Studies of marked birds play a central role in research on avian population dynamics and life histories because they are the main means of estimating survival, recruitment and dispersal. Integrated Population Monitoring aims to combine such information from ringing with data on abundance, productivity and the environment in order to develop population models that address applied questions. Diagnosing the demographic and environmental causes of population declines provides a basis for developing management prescriptions for population recovery. We discuss examples of how such studies are contributing to the conservation of European farmland birds and to understanding declines in Afro-Palaearctic migrants. Adaptive harvest management of North American duck populations demonstrates that similar principles apply to the management of quarry species. Over the next two decades we expect significant advances in our understanding of seasonal variation in demographic rates, of density dependence and of the ways in which continental-scale population dynamics are built up from processes operating at much smaller spatial scales. This will be achieved through analyses of large-scale data sets, the combination of conventional and new technologies and the widespread application of fully integrated modelling methods. Large-scale data gathering through national ringing centres will play a key part in this process.

Many bird populations are changing rapidly as a result of widespread environmental change, and these rates of change are expected to accelerate over coming decades, particularly as a consequence of global climate change (Robinson et al 2009a, Gregory et al 2009). Over the past two decades there has been an increasing conservation focus on declines of relatively widespread and abundant species, in addition to declines in species that are already rare. Indeed, whilst conservationists in Britain and Ireland have had considerable success in reversing declines in very rare species, declines in a range of widespread farmland birds, woodland birds and Afro-Palaearctic migrants are proving much more challenging to address (Eaton et al 2009). To develop effective conservation strategies to reverse such changes, it is necessary to understand the demographic and environmental causes of change, together with the demographic consequences of proposed management actions (Bailie 2001). Where conservation programmes involve introductions of captive-reared or translocated birds, understanding the demographic consequences of these actions is also important.

While the focus of conservation effort tends to be on declining species, similar considerations apply to the management of populations that are hunted and to the control of introduced populations, and other species regarded as pests. These issues of population management require an understanding of the demographic processes and underlying environmental factors that determine abundance. In this paper we review the development and application of techniques for investigating the demographic and environmental processes that bring about changes in bird populations. We give particular emphasis to studies carried out at large spatial scales and to the components of these studies that are based on bird marking. Most such studies are only practical in areas where there are well-established networks of volunteer ringers and birdwatchers. For this reason most such work undertaken to date has been in Europe or North America. There is a very extensive literature on avian population dynamics, life histories and behavioural ecology based on intensive long-term studies of the populations of relatively small study areas (Newton 1998). Such work provides much of the theoretical underpinning for applied studies undertaken at larger spatial scales, but is largely outside the scope of the present review.

PRINCIPLES AND METHODOLOGY

Population dynamics

At its simplest level, population dynamics involves accounting for population changes in terms of births,
deaths, immigration and emigration (Appendix). Where we are dealing with populations at large spatial scales (eg national population indices) it is often reasonable to assume that immigration and emigration will be negligible. Key population dynamics variables can be measured either from intensive local population studies or from extensive programmes involving some combination of censuses, nest recording and ringing.

To understand the causes of population change, it is necessary to measure how individual demographic rates vary in relation to environmental variables and to the abundance of predators, pathogens, and competitors of the same or different species. Competition between individuals of the same species, and sometimes also the impacts of predators and pathogens, give rise to density dependence, which tends to regulate the population around the number of individuals that the environment can support (Newton 1998). Nearly all populations are affected by both density-independent factors (eg severe weather mortality, road deaths) and density-dependent ones (eg competition for food or nest sites). In general it is the relative strength of these factors that determines whether or not the population fluctuates greatly from year to year. However, density-dependent regulation causes the population to fluctuate around a level set by key environmental resources. When these resources change, as might result from a long-term reduction in food supplies or habitat quality, the average population level will be expected to track changing resource levels rather than to remain constant.

The rigour with which the causes of population declines can be diagnosed depends on what information and resources are available (Caughley 1994, Green 1995). In principle it may be possible to relate spatial or temporal variation in abundance directly to environmental resources, without any measurement of underlying vital rates. However, while this approach may work quite well in straightforward cases, it is unlikely to provide robust predictions where more complex population processes are involved. Environmental variables act directly on vital rates and measuring such relationships directly should improve the chances of identifying key variables and quantifying their effects on abundance (Temple & Wiens 1989, Baillie 1990).

Much of our knowledge of the processes that determine avian abundance is derived from correlative analyses of data from a limited number of long-term population studies, most of which have been carried out at small geographical scales (Perrins et al 1991). Such studies are increasingly being complemented by analyses of extensive data gathered at much larger spatial scales, building on the pioneering work of former BTO Directors David Snow (1966), and Raymond O’Connor (1980), which eventually gave rise to the BTO’s Integrated Population Monitoring Programme (Baillie 1990, 2001; Fig 1). The results from such studies should ideally be combined to produce population models that allow the effects of environmental change and management options to be explored, taking account of interactions between key relationships (Potts & Aebischer 1995, Aebischer & Ewald 2004). Where possible, predictions from such field studies and modelling should be tested through experiments. There is now a considerable body of experimental evidence for a range of factors that affect avian abundance and population dynamics (Newton 1998). However, there remain many situations where it is not practical or not ethical to carry out experiments at appropriate spatial scales, making it necessary to rely on evidence from field studies and modelling alone.

Estimating demographic parameters

Survival, recruitment and dispersal rates are normally estimated from studies of marked birds. Two main approaches are used, one based on recoveries of birds that have been killed or found dead and subsequently reported (dead recoveries) and the other on live recaptures or resightings (mark-recapture). There is now a very extensive body of work on these methods (Williams et al 2002, Thomson et al 2009; proceedings of previous EURING analytical conferences listed at www.euring.org). The main parameters in these models are survival rates, reporting rates and recapture rates. These are generally modelled as some function of time, age, location, environmental covariates and population density. Classes of models involving individual covariates (eg body size) are also available. More-complex models may exhibit a wide range of features, including combining both classes of data, movements between subpopulations and transitions between states (eg breeders and non-breeders).

Ring recoveries from dead birds normally are modelled as functions of annual survival rates and ring reporting rates. Most of the information about survival is contained in the recovery data, so higher ring reporting rates lead to more precise estimates of survival for a given amount of ringing effort. Historically, there has been a lack of sufficiently detailed information on numbers ringed, leading to the need to conduct analyses under the unrealistic constraint of constant reporting rates within cohorts. This problem is now being addressed through widespread computerisation of ringing data, but it is still an issue for a number of important historical data sets. There is much evidence of recent declines in reporting rates (eg Robinson et al 2004, Freeman et al 2007, Robinson et al 2009b), which emphasises both the importance of modelling such changes explicitly and the need to improve reporting rates so as to increase the precision of survival estimates. In North America, reward band studies have been used to estimate
the proportions of ringed birds encountered by hunters that are reported to the banding lab. These indicate that without rewards only 32% of the rings encountered by hunters were reported (Nichols et al. 1991). However, the use of a toll-free phone number on bands since 1996 has more than doubled this reporting rate (Tautin et al. 1999).

In the UK, use of a web address on rings has similarly led to a 26% increase in reporting rates and this approach is now being developed throughout Europe.

For the types of studies discussed here, mark–recapture data are normally modelled as functions of survival and recapture rates. A feature of these capture–recapture models is that they cannot distinguish between mortality and permanent emigration from the study area. Thus they provide an estimate of apparent survival which is the joint probability of surviving and not emigrating. Apparent survival is therefore always an underestimation of true survival, but the bias decreases with increasing spatial scale of the study area (Marshall et al. 2004, Schaub et al. 2006). These methods are central to the analysis of data from Constant Effort ringing schemes and the Retrapping Adults for Survival (RAS) project (Robinson et al. 2008, 2009c). A key point arising from such mark–recapture studies is that planned projects with systematic sampling protocols are needed in order to generate high-quality mark–recapture data. For example, the Constant Effort Sites (CES) protocol usually supports models where recapture rates are constant within sites.

Overall recruitment can be estimated from mark–recapture data based on an analysis of reverse capture histories (Pradel 1996). Where a high proportion of breeding adults can be identified as first-year or adult birds (i.e. EURING age codes 5 or 6), recruitment can also be estimated as the proportion of first-year birds in the breeding population (Snow 1966). Given its relatively straightforward nature, this technique has seen surprisingly little use.

Recruitment can be viewed as the product of productivity and first-year survival, while productivity itself is the product of breeding success per attempt, number of attempts and post-fledging survival. Within integrated population models, overall productivity is often estimated by difference, based on a combined analysis of data on abundance and survival (Robinson et al. 2004, King et al. 2008, Baillie et al. 2009a). An index of productivity can be obtained for standardised proportions of juveniles in the population obtained from constant-effort sites (Robinson et al. 2009c) or cannon-netting of waders (Clark et al. 2006).

Age ratios from direct observations in the field can be used to measure the productivity of species such as geese. However, population modelling has shown that, while age ratios of Pink-footed Geese *Anser brachyrhynchus* recorded in Britain in autumn are consistent with other measures of the dynamics of this population, those for Greylag Goose *A. anser* substantially underestimate productivity (Frederiksen et al. 2004). This highlights the importance of validating such data against independent information wherever possible.

Number of fledglings per attempt can be estimated from nest records data and can itself be broken down into its demographic components, usually clutch size, egg survival and nestling or chick survival (Crick et al. 2003). Estimating

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**Figure 1.** Structure of the BTO’s Integrated Population Monitoring Programme (Bailie 1990, 2001). Solid lines indicate the provision of good information for at least some species while broken lines indicate the provision of less-complete information. There is considerable variation in the numbers of species covered by different schemes, and consequently in the numbers of species for which particular vital rates can be estimated.
numbers of attempts per season is more difficult, and usually requires intensive studies based on colour-marked birds. However, this is only a serious problem for multi-brooded species. Indications of the overall length of the breeding season can be obtained from data on brood patches, female weights and moult timing of adults caught for ringing at constant-effort sites or in similar studies.

Survival during the immediate post-fledging survival period tends to be low and is difficult to estimate. Studies based on colour-marked or radio-tagged birds are very labour intensive and may have difficulty separating mortality and emigration, particularly once some juveniles start to disperse a few weeks after fledging (Schmidt et al 2008). Recovery-based studies require large numbers of nestlings, juveniles and adults to be ringed, combined with an adequate reporting rate (Thomson et al 1999, Freeman et al 2007).

**Development of integrated population models**

The initial development of integrated population models for large-scale data involved simple approaches where each vital rate was estimated from a separate analysis, and the demographic parameters were then combined to give a predicted population trajectory. This can then be compared with the population trajectory measured from census data in order to validate the model. In such models some vital rates can be set to their average long-term values, making it possible to determine the extent to which observed variation in individual vital rates is sufficient to explain population declines (Thomson et al 1997, Peach et al 1999). An alternative approach to such analyses is to calculate correlations between population growth rate and each individual demographic parameter. Where data are insufficient to provide annual estimates of each demographic parameter it is possible to calculate estimates for blocks of years identified through an analysis of turning points in the population trajectory (Siriwardena et al 1998, Robinson et al 2004).

More recently, a fully integrated approach to the analysis of such data on population dynamics has been developed. This normally involves the development of a state-space model which incorporates an observation process to account for observation error in the census data and a system process where population changes are a consequence of the underlying demographic rates (Appendix). Using this approach, information about all underlying demographic rates that is included in the population census data can be exploited because it is combined with information about single demographic rates stemming from independent data (eg ring recoveries). Model parameters can then be estimated by maximising the resulting joint likelihoods using the Kalman Filter (Besbeas et al 2002) or by drawing samples from the joint posterior distribution of these likelihoods (and non-informative prior distributions) using Bayesian methods (Brooks et al 2004, Schaub et al 2007). A major advantage of these techniques is that they make the most efficient use of all of the available data under the assumption that all data are samples from the same population. It results in increased precision of parameter estimates which in turn enhances the power to detect relationships between demographic and environmental variables. It also allows variability in the population changes to be quantified (as the observation ‘error’ is partitioned out), which may be important in understanding the adaptability of populations to environmental change (Lande et al 2003). Using Bayesian techniques it is straightforward to derive estimates and credible intervals for correlations between population growth rates and individual vital rates, allowing the contribution of each vital rate to the overall population trend to be evaluated (Brooks et al 2004, Baillie et al 2009a). Integrated modelling methods also allow the support for a wide range of different models to be evaluated, using AIC, BIC or reversible jump MCMC (King et al 2008, Baillie et al 2009a).

One difficulty in identifying the contributions of different demographic rates to population change is that, where a demographic rate is density dependent, changes may be masked by the compensatory effects of the density dependence (Green 1999a). Thus while a density-independent demographic rate responsible for a population decline is likely to show a pronounced long-term change, a density-dependent rate may only show a very brief reduction that would be difficult to detect from monitoring data. Thus there is a need for more attention to be given to models that incorporate density dependence.

To date most such integrated models of population dynamics have focused on the estimation of population growth rates, demographic parameters and the relationships between them. Such models are useful for identifying the demographic causes of previous population changes but they are not well suited to predicting future population levels. Future work should give greater attention to the effects of environmental factors, particularly climate and habitat, and of density dependence, so as to produce models with greater predictive capability.

In developing such models it will be desirable to give greater attention to spatial variation in population dynamics.

Large-scale integrated analyses of population dynamics have tended to focus either on a limited number of specific study sites thought to be representative of a wider population (Peach et al 1995, Grøtan et al 2009) or on extensive data gathered across a very large area such as the whole of Britain (Siriwardena et al 1999, Freeman &...
Crick 2003, Robinson et al 2004). Small-scale studies can be challenging, because immigration into and emigration from the study area have to be modelled in addition to survival and fecundity (Abadi et al unpublished3). Where there are sufficient data, comparisons of population dynamics between regions can reveal large-scale variation in demographic processes (Freeman et al 2007). However, a potentially much more powerful approach is to develop spatio-temporal models of variation in demographic rates (Paradis et al 2000, Royle & Dorazio 2008).

CASE STUDIES

Declines in farmland bird populations

In the UK and throughout much of Europe, declines in farmland bird populations have become a major conservation issue. Some species, particularly Corncrake Crex crex, Stone-curlew Burhinus oedicnemus and Cirl Bunting Emberiza cirlus have small UK populations for which it has been possible to develop effective recovery programmes through carefully targeted agri-environment schemes. Understanding how demography is influenced by agricultural practices has been important in formulating effective management plans for these species (Evans et al 1997, Green & Hirons 1991, Green 1999b, 2008).

Many declines of widespread species started as early as the mid 1970s, but it took a long time for their significance to be recognised. Reversing these changes is much more challenging because it requires modification of agricultural management practices over large parts of the countryside. The first stage in this process has been the successful identification of many of the demographic 12.

The UK Government has adopted a target to reverse the declines of farmland bird species by 2020, based on trends in 20 farmland bird species that make up the farmland bird indicator (Gregory et al 2004). The main delivery mechanism is the Entry Level Stewardship agri-environment scheme (ELS) which provides a broad range of prescriptions suited to the majority of farmland. A novel demographic modelling approach has been used to assess the level of take-up of particular options that is likely to be required to reverse the declines of individual farmland bird species (Siriwardena & Vickery 2003, Vickery et al 2008).

Relationships between the proportion of the population within ELS management and demographic rates were modelled and compared with estimates of the changes in key demographic rates required to increase population size. These models were then combined with estimates of the effects of specific ELS options on key demographic rates to assess the levels of uptake of particular combinations of options that will be needed to achieve population recovery. ELS was predicted to have the desired effect on most of its target species but there were five species for which current management prescriptions require further development. Despite a number of limitations of the models and data (Siriwardena & Vickery 2003, Vickery et al 2008), this work clearly illustrates the applied value of understanding links between management prescriptions, demographic processes and population changes.

Population dynamics of Afro-Palaearctic migrants

It is self evident that the abundance of migratory populations is likely to be affected by processes operating during passage periods and in the wintering areas, as well as on the breeding grounds (Newton 2007). The first key use of ringing data within such studies is to aid identification of the passage and wintering areas of different populations. Such information allows population changes measured on the breeding grounds to be related to environmental conditions in the passage and wintering areas (Baillie & Peach 1992), as well as providing a starting point for population-specific studies of passage and winter ecology. While there are often reasonably good data on the passage areas used by different European populations (Wernham et al 2002, Cepák et al 2008), information on the African wintering grounds is often very sparse. For larger species such information can now be obtained from satellite tracking and there is increased impetus to develop a range of tracking devices for smaller birds (Fiedler 2009) as well as use of intrinsic markers such as stable isotopes and trace elements (Coiffait et al 2009).

A second key use of ringing data is to estimate annual survival rates for breeding populations that can then be correlated with environmental conditions on the wintering grounds. There is a growing body of evidence that the population dynamics of many species wintering in the Sahel region is influenced strongly by winter conditions in that area (Cavé 1983, Peach et al 1991, Baillie & Peach 1992, Eraud et al 2009). For White Storks Ciconia ciconia breeding in eastern Germany and Poland, variation in primary production at one key area in the eastern Sahel visited from October to November explains 88% of annual variation in survival (Schaub et al 2005). Thus for some migratory species it appears that population fluctuations can be driven by conditions during short but critical periods of the annual cycle. While many of the strongest relationships reported to date relate to populations wintering in the Sahel region, survival of species wintering in other areas is also related to weather conditions on the wintering grounds. Data from the RAS scheme show

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3Abadi, F., Gimenez, O., Ullrich, B., Arlettaz, R. & Schaub, M.
Estimation of immigration rate using integrated population models (unpublished manuscript).
that three species of hirundines wintering in different parts of Africa all show relationships with winter weather conditions (Robinson et al. 2008).

However, factors operating on the breeding grounds may also be important, and these relationships may also differ between populations. For example the survival of Hungarian Sand Martins *Riparia riparia* is positively related to Sahel rainfall in their wintering areas (Szép 1995). British Sand Martins also show such relationships but they show a stronger negative relationship between breeding-season rainfall and survival (Cowley & Siriwardena 2005). Increased summer rainfall is thought to reduce the availability of insects on which these birds feed. Ringing will continue to play a key role in future research on Afro-Palaearctic migrants, which needs to focus on ecological processes operating on wintering and passage areas and on building integrated models of the population dynamics of key species.

**Adaptive harvest management**

Effective management for sustainable harvesting of populations that are hunted needs to be based on a good understanding of the population dynamics of the species concerned (Aebischer et al. 1999, Baillie 2001). Autumn and winter hunting inevitably causes an immediate increase in mortality but this may be partially or entirely compensated by changes in density-dependent mortality later in the same season or by density-dependent compensation at other stages of the annual cycle. To obtain a maximum sustainable yield it is generally necessary to maintain a stable population at a level well below that which it would achieve in the absence of harvesting. Thus the management required to maximise harvests may often be different to that required to maximise population size.

Management must usually be based on incomplete knowledge of the population being harvested, and of the most appropriate models for describing its dynamics. Large-scale experiments based on manipulation of hunting regulations would be the most rigorous way to increase knowledge of the dynamics of hunted populations, but this is rarely practical either politically or logistically. Since 1995 a novel approach to the management of hunted populations in the presence of uncertainty has been operated in North America, initially for mid-continent Mallard *Anas platyrhynchos* populations (Nichols et al. 2007). Management decisions are based on a set of (in this case four) competing models of the effects of hunting on harvest and population size. These models have weights which sum to 1 and which represent the relative degree of belief in each model. Each year monitoring data on variables such as population size, harvest, survival and reproduction are collected and used to update current belief in the alternative models via the model weights. The updated models are then used to make predictions that form a basis for setting the next season’s hunting regulations. Thus regulations are based on the most up-to-date understanding of the system, and the effects of varying the hunting regulations feed into increased knowledge of the system. The system is considered to have been successful, particularly in getting managers and ecologists working more closely together and in focusing debates about management options onto the underlying biological processes (Nichols et al. 2007).

A large-scale banding programme is a key element of the monitoring that underpins adaptive harvest management of Mallard in North America, as this provides key estimates of harvest and survival rates. Other monitoring, including breeding counts and hunting-bag surveys, is also required to provide a reasonably complete picture of the dynamics of the system. Overall, the approach is thought to have worked well and is likely to be extended to a range of other waterfowl species (Nichols et al. 2007). Much waterfowl hunting also takes place in Europe, but pan-European monitoring of the vital rates of waterfowl is needed to support more effective harvest management (Elmberg et al. 2006).

The approach used in adaptive harvest management is applicable to a much wider range of conservation and management problems (Williams et al. 2009). For example, in the case of farmland bird recovery programmes discussed above it could be used to develop more effective agri-environment schemes in the context of improving understanding of the effects of large-scale habitat management on the population dynamics of farmland birds.

**FUTURE DIRECTIONS**

A first priority is to continue to build on the successful development of Integrated Population Monitoring and modelling over the last 20 years, maintaining long-term and large-scale studies, enhancing databases and applying the latest analytical techniques. Over the next five to ten years such studies are likely to have particular emphasis on identifying the demographic mechanisms for declines amongst Afro-Palaearctic migrants. Studies to evaluate and predict the effects of global climate change will also be a priority. Demographic monitoring can never be comprehensive and it is important to ensure continuing long-term data collection for key species and that, as far as is practical, these species include a reasonably representative set of ecologies and life histories.

More emphasis should be given to the development of well-designed population studies undertaken at a representative set of sites, building on the success of CES schemes and the BTO’s RAS projects and using new technologies for tracking marked individuals to increase the
range and quality of data (Baillie et al. 2009b, Fiedler 2009).
There should be a continuing effort to improve the quality of
the habitat and environmental data that can be used for
these studies, both by collecting better measurements in the
field and through links to complementary data sets.

Long-term studies have had considerable success in
identifying the main life-cycle stages responsible for
population declines. However, much less work has
focused on identifying the time of year at which the
critical population processes take place. This is particularly
important for migrants, where conservation action should
be guided by knowledge of the way in which critical
population processes are distributed between the breeding,
passage and wintering grounds (Runge & Marra 2005).
Pioneering work on neotropical migrants has shown that,
where long-distance migrants are reasonably site
faithful during the winter, much useful information can
be obtained by conducting mark–recapture studies on the
breeding and wintering grounds (Sillett & Holmes 2002).
Where birds are resident, or populations can be followed
through most of the year, estimation of seasonal survival
rates should be more straightforward (Frederiksen et al
2004) and RAS projects can be structured to provide this
type of information (Robinson et al 2009d). Nevertheless,
such studies remain rare and should be a priority for future
work.

Future integrated population models should adopt a
more process-oriented approach, with greater emphasis
on key environmental variables, seasonal variation in
vital rates and density dependence. This should result in
more-predictive models that can be used to explore the
implications of alternative climate and land-use
scenarios. For many species, developing such models may
be impossible using classical approaches alone, because
long-term times-series of environmental variables may not be
available and density dependence is sometimes difficult to
estimate from long-term demographic data. An alternative
approach to demographic modelling has been to develop
individual-based models (Norris 2004, Goss-Custard &
Stillman 2008). In this approach, behavioural relationships
based on foraging observations and other relevant
behavioural data are used to derive key population-level
relationships such as density-dependent functions, which
are then fed into population models. Both individual-based
models and those based on long-term demographic data
have particular strengths and weaknesses and are best
regarded as complementary approaches. More work is
needed to combine these methods so as to make the best
use of data from both sources. The emergence of Bayesian
techniques for integrated population modelling provides
a natural framework for integrating such techniques at an
analytical level.

Most integrated population analyses to date have
been based on particular study populations or on the
populations of individual regions or countries considered
as a single unit. A natural extension of this approach is
to examine spatial variation in both abundance and the
demographic processes that bring it about (Paradis et al
2000, Royle & Dorazio 2008). Such analyses immediately
offer increased power to explore a range of questions
about population dynamics. However, they also introduce
additional methodological issues, because it is necessary to
account for spatial variation in nuisance parameters such
as detectability and ring reporting rates. Nevertheless we
expect greatly increased emphasis on such spatio-temporal
modelling approaches over the next ten years.

There is much interest in determining how landscapes
can be better designed to support biodiversity, with a
recent emphasis on approaches that could mitigate some
of the predicted effects of global climate change (Opdam
& Wascher 2004). Many populations are profoundly
influenced by habitat structure at the landscape scale
(Hinsley et al. 1996, Freckleton et al. 2005, Dolman et al
2007) and are best viewed as metapopulations whose overall
dynamics are determined by processes operating within
patches and by dispersal between patches. To date, most of
the work in this area has involved modelling relationships
between distributions, abundances and landscape
features. Studies of marked birds can potentially enhance
these approaches by providing direct measurements of
the vital rates of birds living within different classes of
habitat patches and information on movements between
patches (Matthysen et al. 1995, Baillie et al. 2000). New
tracking technologies will increase the extent to which
the use of space by individuals and its consequences
for survival and reproduction can be followed in detail.
This in turn will allow the development of more-refined
models of meta-population dynamics, building on the
approaches described above. Carefully structured projects
coordinated through Ringing Schemes could make a
significant contribution to our understanding of the
dynamics of avian metapopulations at the landscape
scale, and hence to biodiversity conservation within the
wider countryside.

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APPENDIX – ACCOUNTING FOR POPULATION CHANGE

For a typical passerine bird that starts breeding at age one we have:

\[ N_{t+1} = 0.5N_t F_t S_{p,t} + B_t S_{f,t} + N_t S_{a,t} \]

where:

- \( N_t \) is the number of adult females alive in year \( t \)
- \( F_t \) is the number of fledglings per nesting attempt in year \( t \)
- \( B_t \) is the number of nesting attempts per female in year \( t \)
- \( S_{p,t} \) is post-fledging survival (to late summer) in year \( t \)
- \( S_{f,t} \) is first-year survival from years \( t \) to \( t+1 \)
- \( S_{a,t} \) is adult survival from years \( t \) to \( t+1 \)

For simplicity this equation is written in terms of females and assumes a 1:1 sex ratio amongst fledglings. Note that the first set of terms on the right-hand side can potentially be replaced by a single recruitment rate, but with inevitable loss of detail concerning the underlying demographic processes. Such models can readily be extended to more complex matrix models incorporating delayed start of first reproduction and age-, sex- or group-specific variation in fecundity and survival (Caswell 2001).